

Research Paper: Cytotoxic Effects of Bisphenol A as an Endocrine Disruptor on Human Lymphocytes



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ABSTRACT

Background: Endocrine compounds, such as Bisphenol A (BPA), stimulate or inhibit the activities of hormones, nuclear receptors in the central nervous system, liver and other organs. They may be disposed of in the environment inadvertently around industrial sites. The aim of this study was to evaluate the cytotoxic effects of BPA on human lymphocytes in culture at varying concentrations.

Methods: 0.1 mL heparinized 0.2 mL peripheral blood taken from a healthy male and a female were plated in culture media under sterile conditions. To prepare the reference dose at a concentration of 0.05 mg/mL, 0.027g BPA was dissolved in 1 L dimethyl sulfoxide and the highest dose of 50 µg/mL BPA solution was prepared. After separating the stock solution, 50 µg/mL BPA was diluted to prepare 20, 10 or 5 µg/mL doses.

Results: After 24 h of incubation, abnormal cell±Standart Error (%) [AC±SE (%)] 1.10±1.0, chromosomal aberration/cell±Standart Error (CA/cell±SE) 0.025±0.01 was determined in control group, and AC±SE (%) 2.00±0.98 in control group. After 48 h of incubation 0.98, CA/cell±SE was found to be 0.020±0.01. After 24 and 48 h of incubation, AC±SE (%) and CA/cell±SE ratios were 30.00±3.24, 34.00±3.35 and 0.325±0.03, 0.430±0.04, respectively.

Conclusion: The cytotoxic effect of BPA on human lymphocytes was investigated in this study at reference concentration and lower doses. Our findings support the fact that BPA substitutes may not be sufficiently safe for widespread use as industrial chemicals.

Keywords: Bisphenol A, Cytotoxicity, Chromosome aberrations, Endocrine disruptors

Introduction

Endocrine disruptors are synthetic or natural chemical substances that disrupt the normal functioning of body by imitating or inhibiting hormones when consumed [1]. They act by increasing or inhibiting the metabolism of endogenous peptide-rgic or steroidal hormones, or by activating or inhibiting

nuclear receptors in the hypothalamus, adipose tissue, liver and other organs [2]. They may be disposed of in habitable environment as a result of industrial activities or may be found naturally in the environment [3]. These substances are thought to disrupt endocrine balance and affect many metabolic processes, such as growth, stress response, sexual development, reproductive processes, and insulin synthesis [3-5]. In recent years, both the scientific community and the general public have gained

much interest in endocrine disrupting chemicals and their side effects. The basis for the interest is their widespread use, high exposure and possible health problems they may cause. Endocrine disrupting compounds include household chemicals, pesticides, e.g. Dichloro Diphenyl Trichloroethane (DDT), methoxychlor, industrial compounds, heavy metals, such as lead and mercury, flame-extinguishing chemicals (polybromobiphenyls), antibacterial soaps (triclosan), and drug coatings (phthalates) [6, 7]. Further, human exposure may occur occasionally to such natural endocrine disruptors as soybean compounds, like genistein and daidzein [8].

Bisphenol A (BPA) is produced by condensation of two moles of phenol with one mole of acetone at low pH and high temperature, and is generally referred to as 2,2-(4,4-hydroxyphenyl) propane. Bisphenol A comes as solid, crystal structure, in a creamy white color [9]. This compound in 99%-99.8% purity is used primarily as monomers in the production of polycarbonate plastics and epoxy resins [10]. Other uses of BPA are for food storage bags, water and juice bottles, plastic films covering the inner surface of cola and beer cans, beverage bottles, clothing protectors, compact discs, thermal papers, tooth fillers, optical lenses, and baby bottles in a wide range of daily applications [11, 12].

It is believed that people may be exposed to BPA throughout life, starting with intrauterine period [13]. The most common route of exposure is through to be the consumption of food and beverages from polycarbonate bottles and epoxy resin-coated packages [14]. Heating BPA-coated containers at high temperature results in the deterioration of the structure of epoxy resins and release of hydrolyzed BPA products into the foods, primarily due to prolonged use of plastic containers and the long shelf life of such packages [15]. This compound was first evaluated in 1986 by the Scientific Committee for Food (SCF) of the European Union [16]. It was used to evaluate plastic materials and its contact with foodstuff, based on Toxic Dose Index (TDI) being 0.01 mg/kg/day. Since liver detoxification enzymes are not fully developed in fetal and neonatal period yet, it is thought that the toxic effects of BPA are even more serious during this crucial time [17]. According to the report of European Food Safety Authority (EFSA), the No Observed Adverse Effect Level (NOAEL) value of BPA is 5 mg/kg/day and the TDI is 0.05 mg/kg/day [18].

In the Commission's directive #: 2002/72/EC on materials, especially those in contact with foodstuff, the specific migration limit of BPA is set at 0.6 mg/kg [19]. The arrangements made with BPA in Turkey complies with

the European Union directives. The Turkish Food Codex specifies the migration limit of BPA at 0.6 mg/kg in the Communiqué on plastic materials and those in contact with foodstuff [20].

Materials and Methods

Lymphocyte culture: The lymphocytes were taken from two male and female subjects, aged 24 and 25 years, respectively, who were healthy, non-alcoholic, drug-free and without a history of diseases and signed a written informed consent. All procedures were applied according to the guidelines of the International Programme on Chemical Safety (IPCS; a collaborative plan by the United Nations Environment Programme; WHO; Geneva, 1996). The samples were withdrawn as 0.1 mL of heparinized 0.2 mL of peripheral blood, and cultured in 2.5 mL tubes containing culture medium (chromosome medium B) under sterile condition. The culture tubes were then incubated at 37°C for 72 h.

Dose selection: A concentration of 0.05 mg/mL represents the reference dose recommended by EFSA [18]. A 0.01 mg/mL is considered the tolerable daily intake. According to the reference dose, 0.027g of BPA was dissolved in 1L of Dimethyl Sulfoxide (DMSO) to prepare the highest dose of 50 µg/mL BPA solution. Primarily, two third of the prepared solution was used to prepare the 20 µg/mL dose of BPA. This dose was diluted in half to obtain the 10 µg/mL dose, and finally the lowest dose of 5 µg/mL was obtained by diluting the 10 µg/mL dose solution. These doses were added to the medium 24 or 48 h after initiating the culture.

Mitomycin C (MMC) at 0.10 µg/mL was added to the negative and a positive controls and to each dose and chemical groups as well. The colchicine solution was added to each tube at 0.06 µg/mL 2 h before the end of the incubation during the 70 h of the culture period. At the end of the incubation, tubes were centrifuged at 1200 rpm for 10 minutes and the supernatants were discarded. The bottom portion of the tubes (0.5-0.7mL) containing the cells were homogenized by vortexing. Then, the hypotonic solution (0.075 M KCl; 37°C) was added dropwise during vortexing up to 5mL volume.

The tubes were incubated at 37°C for 30 minutes. At the end of incubation, they were centrifuged at 1200 rpm for 10 minutes, and after discarding the supernatants, a pre-cooled fixative solution of 5:1 methanol:acetic acid was added to the tubes dropwise up to 5 mL while being vortexed. These were kept in a refrigerator for 45 minutes. Again, the tubes were centrifuged at 1200 rpm for

10 minutes and supernatant discarded and washed with the fixative solution three times. After the final washing with the fixative, the precipitate, consisting of 0.5-0.7 mL of white blood cells remained at the bottom of tubes, which were homogenized by pipetting. The suspension, which was drawn into a pasteur pipette, was previously cleaned in 1 N nitric acid (HNO₃) and the cells were spread in different areas of humid slides in a refrigerator in 70% ethyl alcohol, and the chromosomes were spread. These preparations were allowed to dry for 24 h at room temperature in the dark. The coverslips were then placed on the slides with entellan.

Chromosome aberration: Chromosome Aberration (CA) is one of the most commonly used methods in genetic studies to evaluate the effects of known or suspected genotoxic substances on chromosomes. It is used to determine the frequency of chromosomal abnormalities from peripheral blood lymphocytes in culture.

Statistical analyses: In this study, regression analysis was performed, using SPSS v 24 software to reveal the dose-response relationship in chromosome abnormality study of the cells. The z distribution test was used to deter-

mine whether chromosomal abnormalities in the experimental groups differed from those in the control group.

Results

The doses administered and the observed chromosomal abnormalities are shown in Table 1. After 24 h of incubation, AC±SE (%) 1.10±1.00, CA/cell±SE 0.025±0.01 were determined in the control group. Also, AC±SE (%) 2.00±0.98 was measured in the control group after 48hr of incubation. The CA 0.98, CA/cell±SE was found to be 0.020±0.01. After 24 and 48 h of incubation, AC±SE (%) and CA/cell±SE ratios were 30.00±3.24, 34.00±3.35 and 0.32±0.03, 0.430±0.04, respectively. After 24 h of incubation, 3 chromosome fractures, 25 chromatite fractures, 3 polyploidy, 7 disenteric chromosomes, 18 sister chromatite junction and 12 chromatic changes were detected in MMC medium. After the 48 h incubation, 7 chromatite fractures, 3 polyploidy, 6 disenteric chromosomes, 42 sister chromatite junction and 31 chromatite changes were observed. These chromosomal abnormalities in MMC culture media were found to be statistically significant as compared to those in the control group (P<0.05). No chromosomal abnormality was observed as a result of

Table 1. Observed chromosomal abnormalities

Practice			Chromosomal Abnormalities									
Test Article	Duration (h)	Conc'tion (ppm)	chf	cf	f	p	dic	scj	cc	AC±SE (%)	CA/Cell±SE	
Control	24	0	-	-	-	-	-	2	-	1.10±1.00	0.025±0.01	
MMC	24	0.20	3	25	-	3	7	18	12	30.00±3.24	0.325±0.03	
		5	-	-	-	-	-	-	-	Toxic	Toxic	
		10.0	-	-	-	-	-	-	-	-	Toxic	Toxic
		20.0	-	-	-	-	-	-	-	-	Toxic	Toxic
		50.0	-	-	-	-	-	-	-	-	Toxic	Toxic
Control	48	0	-	1	-	-	-	3	-	2.00±0.98	0.020±0.01	
MMC	48	0.20	-	7	-	3	6	42	31	34.00±3.35	0.430±0.04	
		5	-	-	-	-	-	-	-	-	Toxic	Toxic
		10	-	-	-	-	-	-	-	-	Toxic	Toxic
		20	-	-	-	-	-	-	-	-	Toxic	Toxic
		50	-	-	-	-	-	-	-	-	Toxic	Toxic
BPA	24	0.20	-	-	-	-	-	-	-	-	-	
		5	-	-	-	-	-	-	-	-	-	
		10.0	-	-	-	-	-	-	-	-	-	
		20.0	-	-	-	-	-	-	-	-	-	
		50.0	-	-	-	-	-	-	-	-	-	
BPA	48	0.20	-	-	-	-	-	-	-	-	-	
		5	-	-	-	-	-	-	-	-	-	
		10	-	-	-	-	-	-	-	-	-	
		20	-	-	-	-	-	-	-	-	-	
		50	-	-	-	-	-	-	-	-	-	

h: hour; ppm: parts per million; Conc'tion: Concentration; chf: chromosome fractures; cf: chromatite fractures; f: fragment; p: polyploidy; dic: dicentric chromosome; scu: sister chromatid junction; cc: chromatid changes; MMC: Mitomycin C; AC: Abnormal Cell; SE: Standard Error; CA: Chromosomal Aberration; BPA: Bisphenol A

bursting the cells at all doses of 5, 10, 20 or 50 $\mu\text{g}/\text{mL}$ after either 24 or 48 h of incubation (Table 1).

Discussion

In recent years, interest in endocrine disrupting chemicals has increased. Endocrine disruptors are frequently used in many industries because of their low production costs and their utility in a wide range of applications. Bisphenol A is an endocrine disruptor and one of the commonly used chemicals. Although the toxicity of BPA has been demonstrated by *in vitro* and *in vivo* studies, the previous findings are still inconclusive. Recently, studies on CA in which the effects of BPA on human lymphocytes were evaluated, were taken as reference due to its acceptance by EFSA (0.05mg/mL) [18]. The frequency of CA in the lymphocytes from human peripheral blood has been directly associated with increased risk of cancer and carcinogenesis [21-23]. In contrast, no cytogenetic effect has been reported, according to TDI established by European Union (EU) (2002) at a concentration of 0.01 mg/mL. This indicates that the above dose is safer for human health than the previously used dose at 0.05 mg/mL.

In this study, it was demonstrated that the doses ≤ 0.05 mg/mL, which were accepted as reference by EFSA, still showed cytotoxic effects. In a previous study using five different doses added to lymphocyte cultures, the ones at 0.20, 0.10 and 0.05 $\mu\text{g}/\text{mL}$ caused chromosomal damage. However, the 0.01 $\mu\text{g}/\text{mL}$ dose showed no cytotoxic effect [24]. In another study, BPA doses of 0.4, 1, 4, 40, and 100 $\mu\text{g}/\text{mL}$ were reported to be cytotoxic, although differing results were provided [25]. In a Chinese study, the relationship between BPA at varying concentrations in the urine and a history of recurrent abortions were investigated among 102 women with recurrent abortions compared to 162 controls [26]. This study reported that the BPA concentrations in the urine of women with recurrent abortion were higher than that in the control group. Specifically, the increased urine BPA concentrations and a history of recurrent abortion were found to be positively correlated [26].

Conclusions

Bisphenol A, being a lipophilic compound, is migrated into foods with high heat treatment. There is also BPA exposure through drinking contaminated water and dermal contact with BPA exposed air and soil. Although BPA is largely detoxified in the liver when consumed orally, its metabolism on the contaminated skin is unknown. In addition, studies have detected BPA in blood, urine, placenta, breast milk and various tissues and organs. The

toxic effects of BPA are fairly known when it enters the body mixed with foods. The cytotoxic effects of BPA on human lymphocyte have been demonstrated by this study for the first time, in which its effect was evaluated at reference and lower doses accepted by EFSA. Our findings support the conclusion that most BPA products might not be as safe as previously believed.

Ethical Considerations

Compliance with ethical guidelines

All procedures in this study were performed after obtaining voluntary and signed consent forms from the participants. Also, we followed the legal and regulatory requirements for human experimentation.

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Author's contributions

Conceptualization: Mustafa Özgür and Aslı Uçar; Methodology: Mustafa Özgür and Serkan Yılmaz; Investigation: Mustafa Özgür, Şemsi Gül Yılmaz, and Serkan Yılmaz; Writing – Original Draft: Mustafa Özgür, Aslı Uçar; Writing – Review & Editing, Funding Acquisition, All Authors; Resources: Mustafa Özgür and Şemsi Gül Yılmaz; Supervision: Aslı Uçar and Serkan Yılmaz.

Conflict of interest

The authors declare no conflict of interest with any internal or external entity in conducting this study.

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References

- [1] Solecki R, Kortenkamp A, Bergman Å, Chahoud I, Degen GH, Dietrich D, et al. Scientific principles for the identification of endocrine-disrupting chemicals: A consensus statement. *Arch Toxicol.* 2017; 91(2):1001-6. [DOI:10.1007/s00204-016-1866-9] [PMID] [PMCID]
- [2] Bilal M, Asgher M, Iqbal HM, Hu H, Zhang X. Bio-based degradation of emerging endocrine disrupting and dye-based pollutants using cross-linked enzyme aggregates. *Environ Sci*

- Pollut Res Int. 2017; 24(8):7035-41. [DOI:10.1007/s11356-017-8369-y] [PMID]
- [3] Kabir ER, Rahman MS, Rahman I. A review on endocrine disruptors and their possible impacts on human health. *Environ Toxicol Pharmacol*. 2015; 40(1):241-58. [DOI:10.1016/j.etap.2015.06.009] [PMID]
- [4] Jambor T, Tvrđá E, Tušimová E, Kováčik A, Bistáková J, Forgács Z, et al. In vitro effect of 4-nonylphenol on Human Chorionic Gonadotropin (HCG) stimulated hormone secretion, cell viability and reactive oxygen species generation in mice Leydig cells. *Environ Pollut*. 2017; 222:219-25. [DOI:10.1016/j.envpol.2016.12.053] [PMID]
- [5] Sifakis S, Androutsopoulos VP, Tsatsakis AM, Spandidos DA. Human exposure to endocrine disrupting chemicals: Effects on the male and female reproductive systems. *Environ Toxicol Pharmacol*. 2017; 51:56-70. [DOI:10.1016/j.etap.2017.02.024] [PMID]
- [6] Medical Research Council. IEH assessment on environmental oestrogens: Consequences to human health and wildlife. Leicester: Institute for Environment & Health, University of Leicester; 1995. https://www.google.com/books/edition/IEH_Assessment_on_Environmental_Oestrogen/mfwJAQAAMAA?hl=en
- [7] Barrios-Estrada C, de Jesús Rostro-Alanis M, Muñoz-Gutiérrez BD, Iqbal HMN, Kannan S, Parra-Saldívar R. Emergent contaminants: Endocrine disruptors and their laccase-assisted degradation-a review. *Sci Total Environ*. 2018; 612:1516-31. [DOI:10.1016/j.scitotenv.2017.09.013] [PMID]
- [8] Bolt HM, Degen GH. Comparative assessment of endocrine modulators with oestrogenic activity. II. Persistent organochlorine pollutants. *Arch Toxicol*. 2002; 76(4):187-93. [DOI:10.1007/s00204-002-0336-8] [PMID]
- [9] Staples CA, Dome PB, Klecka GM, Oblock ST, Harris LR. A review of the environmental fate, effects, and exposures of bisphenol A. *Chemosphere*. 1998; 36(10):2149-73. [DOI:10.1016/S0045-6535(97)10133-3]
- [10] Palladino G, Sereni L. Bisphenol A in Chronic Kidney Disease. In: Bisphenol A exposure and health risks. In: Erkekoglu P, Belma Koçer-Gümüşel B, editors. *Bisphenol A: Exposure and Health Risks*. London: IntechOpen; 2017. [DOI:10.5772/intechopen.68681]
- [11] Markey CM, Rubin BS, Soto AM, Sonnenschein C. Endocrine disruptors: From wingspread to environmental developmental biology. *J Steroid Biochem Mol Biol*. 2002; 83(1-5):235-44. [DOI:10.1016/S0960-0760(02)00272-8]
- [12] vom Saal FS, Hughes C. An extensive new literature concerning low-dose effects of bisphenol A shows the need for a new risk assessment. *Environ Health Perspect*. 2005; 113(8):926-33. [DOI:10.1289/ehp.7713] [PMID] [PMCID]
- [13] Vom Saal FS. Triennial reproduction symposium: Environmental programming of reproduction during fetal life: Effects of intrauterine position and the endocrine disrupting chemical bisphenol A. *J Anim Sci*. 2016; 94(7):2722-36. [DOI:10.2527/jas.2015-0211] [PMID]
- [14] Michalowicz J. Bisphenol A-sources, toxicity and biotransformation. *Environ Toxicol Pharmacol*. 2014; 37(2):738-58. [DOI:10.1016/j.etap.2014.02.003] [PMID]
- [15] Schönfelder G, Wittfoht W, Hopp H, Talsness CE, Paul M, Chahoud I. Parent bisphenol A accumulation in the human maternal-fetal-placental unit. *Environ Health Perspect*. 2002; 110(11):703-7. [DOI:10.1289/ehp.110-1241091] [PMID] [PMCID]
- [16] Poole A, van Herwijnen P, Weideli H, Thomas MC, Ransbotyn G, Vance C. Review of the toxicology, human exposure and safety assessment for bisphenol A diglycidylether (BADGE). *Food Addit Contam*. 2004; 21(9):905-19. [DOI:10.1080/02652030400007294] [PMID]
- [17] Matsumoto J, Yokota H, Yuasa A. Developmental increases in rat hepatic microsomal UDP-glucuronosyltransferase activities toward xenoestrogens and decreases during pregnancy. *Environ Health Perspect*. 2002; 110(2):193-6 [DOI:10.1289/ehp.02110193] [PMID] [PMCID]
- [18] European Food Safety Authority. Opinion of the Scientific Panel on food additives, flavourings, processing aids and materials in contact with food (AFC) related to 2,2-BIS(4-Hydroxyphenyl) Propane. *Eur Food Saf Author*. 2007; 5(1):1-5. [DOI:10.2903/j.efsa.2007.428]
- [19] Commission Directive. Relating to plastic materials and articles intended to come into contact with foodstuffs [Internet]. 2002 [Updated October 2009]. Available from: <https://eur-lex.europa.eu/LexUriServ/LexUriServ.do?uri=CONSLEG.2002L0072:20091109:EN:PDF>.
- [20] Takeuchi T, Tsutsumi O, Ikezaki Y, Takai Y, Taketani. Positive relationship between androgen and the endocrine disruptor, bisphenol A, in normal women and women with ovarian dysfunction. *Endocr J*. 2004; 51(2):165-9. [DOI:10.1507/endocrj.51.165] [PMID]
- [21] Bonassi S, Znaor A, Norppa H, Hagmar L. Chromosomal aberrations and risk of cancer in humans: An epidemiologic perspective. *Cytogenet Genome Res*. 2004; 104(1-4):376-82. [DOI:10.1159/000077519] [PMID]
- [22] Bonassi S, El-Zein R, Bolognesi C, Fenech M. Micronuclei frequency in peripheral blood lymphocytes and cancer risk: Evidence from human studies. *Mutagenesis*. 2011; 26(1):93-100. [DOI:10.1093/mutage/geq075] [PMID]
- [23] Garcia-Sagredo JM. Fifty years of cytogenetics: A parallel view of the evolution of cytogenetics and genotoxicology. *Biochim Biophys Acta*. 2008; 1779(6-7):363-75. [DOI:10.1016/j.bbagr.2008.05.003] [PMID]
- [24] Santovito A, Cannarsa E, Schleicherova D, Cervella P. Clastogenic effects of bisphenol A on human cultured lymphocytes. *Hum Exp Toxicol*. 2018; 37(1):69-77. [DOI:10.1177/0960327117693069] [PMID]
- [25] Aghajani-pour-Mir SM, Zabihi E, Akhavan-Niaki H, Keyhani E, Bagherizadeh I, Biglari S, et al. The genotoxic and cytotoxic effects of Bisphenol-A (BPA) in MCF-7 cell line and amniocytes. *Int J Mol Cell Med*. 2016; 5(1):19-29. [PMID] [PMCID]
- [26] Shen Y, Zheng Y, Jiang J, Liu Y, Luo X, Shen Z, et al. Higher urinary bisphenol A concentration is associated with unexplained recurrent miscarriage risk: Evidence from a case-control study in eastern China. *PLoS One*. 2015; 10(5):e0127886. [DOI:10.1371/journal.pone.0127886] [PMID] [PMCID]

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